

Risk Assessment of non-listed substances (NLS) and non-intentionally added substances (NIAS) under Article 19 of Commission Regulation (EU) No 10/2011 of 14 January 2011 on plastic materials and articles intended to come into contact with food

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Risk Assessment of non-listed substances (NLS) and non-intentionally added substances (NIAS) under Article 19 of Commission Regulation (EU) No 10/2011 of 14 January 2011 on plastic materials and articles intended to come into contact with food¹

With this guidance document, PlasticsEurope intends to explain how the plastics (plastic intermediate material) producers interpret and respond to their risk assessment obligations for non-listed and non-intentionally added substances under the article 3 of the Framework Regulation (Regulation (EC) No 1935/2004) and article 19 of the Regulation on plastic materials and articles intended to come into contact with food (Regulation (EU) No 10/2011), based on internationally recognized tools and scientific knowledge available to them at the time of writing.

This is a living document which will be updated when and if needed.

1 Risk-based legislation

The EU food contact regulation is a risk-based regulation. This means that the regulation is not based on the hazard potential of substances in plastic materials for food contact but on the risk assessment that has been carried out by the EU risk assessing agency (EFSA, European Food Safety Authority) for mainly additives and monomers. A risk assessment consists of three components: hazard identification and characterisation, appraisal of exposure, followed by the risk assessment itself, which is then 'translated' by the EU legislative body (here DG Health and Consumer Protection) by setting migration limits for substances from plastics into the food.

'Hazard' is the potential of something to cause harm. Hazard typically refers to the intrinsic properties of a chemical, such as toxicity, while 'exposure' addresses the likelihood and degree to which a human or environmental receptor will be exposed to the intrinsic hazards of a chemical. 'Risk' is the likelihood of harm occurring.

Captured into a simple formula, this would read: hazard x exposure potential = risk

'Risk assessment' puts hazard and exposure together in an attempt to understand the 'real world danger' posed by a chemical based on its intrinsic hazards in the light of anticipated exposure.



¹ <http://eur-lex.europa.eu/LexUriServ/LexUriServ.do?uri=CONSLEG:2011R0010:20111230:EN:PDF>



For a hazardous object or situation to become a risk, there must be exposure. For example, a wild and dangerous animal will always represent a hazard, but as long as it remains properly caged it will never represent a risk. As exposure increases, so does the likelihood of harm, and therefore the risk will be higher.

Risk assessments are scientific analyses to determine the level of exposure to a hazard and attempt to give accurate levels of potential harm to human health and the environment.



Exposure assessment: no unique scenario

Risk assessment certainly requires the rigorous and objective analysis of data. However, there may be weaknesses or gaps in the data that can only be addressed by applying professional judgment. The various assumptions and uncertainties carried over from the hazard characterisation and exposure assessments also affect the risk assessment. Extrapolating from that work to reach probabilistic conclusions necessarily creates additional uncertainties²

Principles of Article 19

To ensure compliance with article 3 of the Framework Regulation (EC) No. 1935/2004, article 19 allows industry to conduct self-assessments following the internationally recognised scientific principles for risk assessment of substances used in plastic materials and articles, exempt from authorisation in Commission Regulation (EU) No. 10/2011: colorants, solvents, polymeric production aids (PPAs), aids to polymerisation (APs) and non-intentionally added substances (NIAS).

For those non-listed substances which do not need to be authorized, the following methodology could be used:

- Verify if the substance is authorized by international recommendations or at national level or
- Do a risk assessment on the basis of internationally recognised scientific principles according to Article 19.

What does this paper intend to deliver?

² Hazard v. Risk in EU Chemicals Regulation, *Kristina Nordlander, Carl-Michael Simon and Hazel Pearson – EJRR 3/2010*

This Guideline provides an overview of approaches for assessing the risks of non-listed substances as required by Article 19 of Commission Regulation (EU) No 10/2011 on Plastic Materials and Articles intended to come into contact with food, in this guideline referred to as the Regulation. The Regulation contains in Annex I a list of monomers, starting substances and additives evaluated and authorized for use in food contact plastics (i.e., the Union list).

Under the provisions of the Regulation, a number of substances present in food contact plastics are exempt from the requirement to be included in the Community positive list (Union list) according to Article 6. The substances exempted of positive listing include: solvents, colorants, polymer production aids (PPA's), aids to polymerization (AP's), oligomers and the so-called "non-intentionally-added-substances (NIAS)". NIAS include substances such as impurities, contaminants, reaction, decomposition or degradation products. It should therefore be made clear that this guideline document applies to substances exempted from authorization at the EU level according to the Article 19, unless already listed and subject to restrictions.

The non-listed substances are subject to the provisions of Article 3 of the Framework Regulation (EC) No 1935/2004 that applies to all food contact materials. Article 3 states that exposure to substances from food contact materials should not pose a risk to human health. For non-listed substances this should be demonstrated through a risk assessment and documented in the internal supporting documentation package for supporting the Declaration of Compliance. Article 19 of the Regulation articulates the need for a risk assessment for non-listed substances in accordance with internationally recognized scientific principles on risk assessment. This paper provides an overview on how to perform a risk assessment according to internationally recognized scientific principles on risk assessment as asked for by the Regulation. Note that extensive literature is available on risk assessment methodology and that slightly different approaches and terminology are used by different organisations. A risk assessment according to internationally recognized scientific principles as asked for by the regulation will be carried out by regulatory experts familiar with the different approaches and their bases. The non-exhaustive list of references at the end of this chapter provides further sources of information on risk assessment (guidance documents on regulatory websites^{2, 3, 4, 5}, scientific publications, reports, etc...).

The Packaging Materials Task Force at ILSI-Europe³ has published a range of useful technical reports on risk assessment for food packaging⁶. The website of the US Food and Drug Administration also provides useful guidance on risk assessments^{4, 5, and 7}.

2 A risk assessment typically consists of the following four components:

- 1. Literature survey on existing legislation**
- 2. Exposure assessment**
- 3. Toxicological assessment**
 - Hazard identification
 - Dose-response assessment (Hazard Characterisation)
- 4. Risk characterisation**

³ International Life Science Institute, a non-profit worldwide foundation established to advance the understanding of scientific issues related to nutrition, food safety, toxicology and the environment

2.1 Literature survey on existing legislation

It is general practice to first check existing legislation.

Most of this can be carried out in-house or via the web on public or specialised websites.



2.2 Exposure Assessment

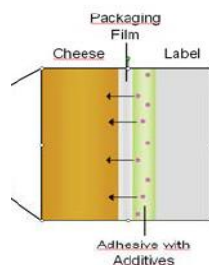
Exposure assessment aims to define the dose of non-listed substances that individuals receive in exposed populations. This dose is the so-called Estimated Daily Intake (EDI) (mg/person/day).

The EDIs for non-listed substances in food contact materials are estimated in a number of ways depending on the material and the nature of the contact. To assess the dietary exposure to a substance migrating from “repeated-use” applications (e.g. pipes, tubing, food containers, and food processing equipment) conservative models are applied. The assessment of food contact materials that come in contact with foods in non-repeated use applications e.g. food packaging is more complex and often require more refined models and additional data. In both cases, tiered approaches are typically used in exposure assessments. Tiered approaches begin by using simple, conservative, and widely applicable models of exposure. These models require relatively little data but tend to overestimate exposures. If the exposure estimates are found to be too large using the conservative models then the assessor moves on to more refined methods.

All exposure assessments for non-listed substances require the same types of information. These include data on the ability of the substance to migrate from the material into food or water during contact events (migration data), and data that allow the prediction of the daily dose to exposed individuals (food consumption and food packing data). The findings of migration are a property of the substance, the food contact material, the food, the duration and conditions of the contact (Temperature, S/V). The findings of exposure are determined by how much food and water are consumed by the average consumer and what types and shapes of packaging are used for the food and water.

2.2.1 *Migration data*

The migration levels of substances from plastics into the food under the typical conditions of use can be derived from worst case calculations (modeling assuming 100% migration), migration calculation models (diffusion model) or migration studies in food simulants (experimental data). Migration is typically expressed in mg/dm² plastic or mg/kg food. For applications where the material or article is intended to come into repeated contact with foodstuffs (so-called “repeated-use” applications e.g. pipes, tubing, food containers, food processing equipment) the migration level in the third migration test should be taken as basis for the risk assessment (Annex V chapter 3.3 of Regulation (EU) No. 10/2011).



2.2.1.1 **Worst-case migration calculation (general):**

The following two formulae can be used to calculate the worst case concentration levels in the food assuming 100% migration.

$$C_{Food} (mg/kg) = \frac{C_{Polymer} (mg/kg) * d_{Polymer} (g/cm^3) * S_{Packaging} (cm^2) * e_{Packaging} (cm)}{M_{Food} (g)}$$

$$C_{Food} (mg/kg) = \frac{C_{Polymer} (mg/kg) * d_{Polymer} (g/cm^3) * S_{Packaging} (cm^2) * e_{Packaging} (cm)}{d_{Food} (g/cm^3) * V_{Food} (cm^3)}$$

With

- $C_{Polymer}$: concentration of the substance in the polymer
- C_{Food} : concentration of the substance into the food
- $d_{Polymer}$: density of the polymer
- d_{Food} : density of the food
- $e_{Packaging}$: thickness of the packaging material
- $S_{Packaging}$: contact area of the packaging material
- V_{Food} : volume of the food in contact with the material
- M_{Food} : weight of the food in contact with the material



By convention and as documented in the EFSA Note for Guidance^{3(p 91)}, it is assumed that for most plastics, migration under typical conditions of use primarily takes place from the first 250 microns of the plastic layer in contact with the food with the exception of plasticized polymers and of the migration of components with low diffusion coefficients (volatile components).

The rate of migration is a function of the substance, the plastic, the food, the contact time and temperature of the food. Data on migration rates are generated based on empirical studies of specific plastics. Such studies typically investigate different types of foods. Ideally, data should be developed for the following types of food: aqueous food, alcoholic food, acidic food and fatty food.

2.2.1.2 Worst-case migration calculation for pipes (repeated use, dynamic state):

The initial Mass concentration C_0 of the substance is assumed to be uniform into the polymer.

The initial mass concentration of the substance in the polymer is:

$$C_{Pol,0} (mg / kg) = \frac{M_{Sub} (g)}{M_{Pol} (g)} = \frac{M_{Sub} (g)}{d_{Pol} (g/cm^3) * V_{Pol} (g/cm^3)}$$

Geometry of the pipe:

- D_{ext} (cm): external diameter of the pipe
- D_{int} (cm): internal diameter of the pipe
- E (cm): thickness of the pipe
- L (cm): length of the pipe
- e (cm): wetted thickness of the polymer
- V_{Food} (cm³): volume of the food in the pipe
- $d_{Polymer}$ (g/cm³): density of the polymer
- d_{Food} (g/cm³): density of the food
- V_{Pol} (cm³): volume of the polymer constituting the pipe



$$D_{ext} = D_{int} + 2 * E$$

$$V_{Pol} = V_{ext} - V_{int} = \pi L E (d_{int} + E)$$

$$V_{Food} = \frac{\pi}{4} L d_{int}^2$$

Assumption:

100% migration of the substance into the foodstuffs of each cycle.

$$C_{Food} (mg/kg) = \frac{C_{Pol} (mg/kg) \cdot d_{Pol} (g/cm^3) \cdot V_{Pol} (cm^3)}{M_{Food} (g)} = \frac{C_{Pol} (mg/kg) \cdot d_{Pol} (g/cm^3) \cdot V_{Pol} (cm^3)}{d_{Food} (g/cm^3) \cdot V_{Food} (cm^3)}$$

$$C_{Food} (mg / kg) = 4 C_{Pol} (mg / kg) \frac{d_{Pol} (g / cm^3) E (cm) (D_{int} (cm) + E (cm))}{d_{Food} (g / cm^3) D_{int}^2 (cm^2)}$$

The concentration of a substance migrating from a pipe is independent of the length of the pipe. Only the fraction of the substance contained in the “wetted” thickness e of the low diffusivity polymer could migrate. Consequently in the above equation the thickness E of the pipe can be replaced by the “Wetted” thickness e (for drinking water the wetted substance is equal to 100 µm according to the French circular DGS-VS4 n°99-217 published the 12 April 1999. For foodstuffs the wetted thickness is 250 µm according to the EFSA note of Guidance-page 55).

After a first migration, the new concentration of the substance is considered as homogeneous in the entire polymer. So the concentration is:

$$C_{Pol,1} = C_{Pol,0} - C_{Food,0}$$

After n cycles of migration, the concentration of the substance in the polymer is:

$$C_{Pol,n} = C_{Pol,0} - \sum_{p=1}^n C_{Food,p}$$

After n cycles of use, the amount of food in contact with the polymer is:

$$M_{Food} (g) = d_{Food} \cdot V_{Food} = n \cdot d_{Food} \cdot \frac{\pi}{4} L D_{int}^2$$

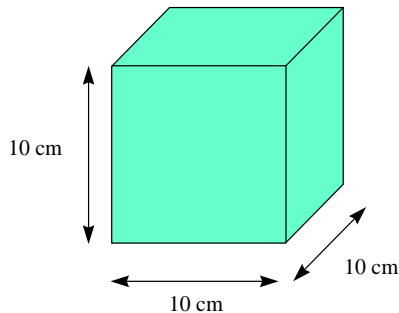
The average concentration of the substance in the food can be calculated assuming 100% of migration into the total amount of food in contact with the polymer during the lifetime of the pipe. This is the total amount of the substance distributed throughout the total amount of food transported during lifetime of the pipe.

$$\bar{C}_{Food} (mg / kg) = \frac{C_{Pol,0} V_{Pol} d_{pol}}{n d_{Food} LD_{int} \frac{2 \pi}{4}}$$

2.2.2 Food consumption and packaging use data

The default (worst-case) assumption in Europe is that every day an adult person consumes 1 kg of food packaged in a 1 dm³ cube with a surface area of 6 dm². It is assumed that the cube is covered by a single type of the same food contact material and the food is the most aggressive extractor of the substance. The individual has the same exposure every day throughout his life. This assumption is the basis for a default food contact rate for the material of 6 dm²/person/day.

This is a conservative assumption that does not reflect any real consumption pattern. In reality only a certain percentage of the daily consumed food is packaged in any one food contact material and within plastics a certain percentage is used to package aqueous food, acidic food, alcoholic food, and fatty food.



EU food packaging cube

Total volume = 1 dm³

Total surface = 6 dm²

2.2.3 Derivation Estimated Daily Intake number – three steps approach

In order to calculate the estimated exposure, a tiered, three steps approach is suggested starting with a simple worst case calculation up to very accurate estimates using highly sophisticated probabilistic assessment models. Such refinements require additional information and are more complex and resource intensive than the conservative approach.

Step 1: Worst case exposure calculation based on European default assumption

Based on the default assumption in Europe that every day an adult person consumes 1 kg of food packaged in a 1 dm³ cube, an estimated worst-case daily intake number can be calculated using the following simple formula.

$$EDI_{\text{worst case}} (\text{mg/person/day}) = 1 \text{ kg food/person/day} * \text{Migration (mg/kg food)}$$

Step 2: FRF corrected exposure calculation for lipophilic substances in fatty food

To account for the fact that 95% of the population consumes less than 200 g of fat per day, the Regulation facilitates dividing the migration levels of lipophilic substances into foods containing more than 20% fat, with a Fat Reduction Factor (FRF). The FRF may vary from 1 to 5 (FRF=1 for food

with a fat content of 20% and FRF=5 for food with fat content of 100%). The FRF corrected exposure number can be calculated using the following formula.

$$EDI_{FRF \text{ corrected}} \text{ (mg/person/day)} = \frac{1 \text{ kg food/person/day} * \text{Migration (mg/kg food)}}{FRF_{\text{lipophilic substances}}}$$

Annex V chapter 4.1 of the Regulation (EU) No. 10/2011 describes the application of the FRF and the specific cases where the FRF cannot be applied (e.g. infant food). The application of the FRF shall not lead to a specific migration exceeding the overall migration limit.

Step 3: Refined exposure calculation using food distribution/consumption factors

A. FDA exposure assessment

The US Food and Drug Administration (FDA) have generated consumption data for packaged food which are used in risk assessments for regulatory purposes. The food consumption data for different packaging materials (Consumption Factors and Food-Type Distribution Factors) are accessible on the FDA website together with detailed guidance on how to calculate consumer exposure to substances migrating from packaging materials⁵.

Under FDA approach, the term "Consumption Factor" (CF) is used to describe the fraction of the daily diet expected to contact specific packaging materials. The CF represents the fraction of daily consumed packaged food that is packed in a certain packaging material. FDA assumes that an individual consumes 3 kg of packaged food (1.5 kg solid and 1.5 kg liquid) per day of which about 80% is packaged in plastics. The Food-type distribution factors (f_T) reflect for each packaging material the fractions of all food contacting each material that is aqueous, acidic, alcoholic and fatty. These data (consumption factors and food type distribution factors) are available on the FDA website (see link):

<http://www.fda.gov/Food/default.htm> (Documents UCM081825 and UCM081818)

The Estimated Daily Intake (EDI) (mg/person/day) of a substance migrating from a plastic packaging material into a specific type of food can be calculated using the following formula (calculated for a 60 kg person and an intake of 3 kg of packaged food per day):

$$EDI_{FDA} \text{ (mg/person/day)} = 3 \text{ kg food/person/day} * CF * <M> \text{ (mg/kg food)}$$

- EDI: Estimated Daily Intake (mg/person/day)
- CF: Consumption Factor for the particular plastic
- <M>: Migration level of substance from the plastic into the food (mg/kg food)

In case specific migration levels are available for the different types of food then the formula can be further refined as follows:

$$<M> = f_{\text{aqueous and acidic}} \cdot (M_{10\% \text{ ethanol}}) + f_{\text{alcohol}} \cdot (M_{50\% \text{ ethanol}}) + f_{\text{fatty}} \cdot (M_{\text{fatty}})$$

f: Food-Type distribution factors for the particular plastic

M: Migration levels of the substance measured in different types of food simulants

The website of US FDA contains a database with cumulative estimated daily intakes (CEDIs) and acceptable daily intakes (ADIs) for a large number of food contact substances⁷.

B. European consumer exposure assessment tools

B1. Risk Assessment of non-intentionally added substances (NIAS) using the MATRIX

Only limited food consumption data are publicly available for plastic packaging materials in Europe^{6f}. The assessment of NIAS apart from their occurrence also requires exposure data for the specific plastics materials.

The Matrix Project was jointly initiated, financed and supported by Cefic-FCA, European Plastics Converters (EuPC), Flexible Packaging Europe (FPE) and PlasticsEurope⁸. Within the project generic levels of migration into food for respective packaging plastics materials were derived. **Above these levels every migrant should be identified and assessed**, however, below which the corresponding exposure is so minor that further assessment could be neglected. This level has been defined "Level of Interest (LOI)": it is linked to each packaging material and will be a function of the exposure of consumers to this material. The calculation of the LOI follows similar conditions as applied to non-listed substances used behind a functional barrier as described in the articles of regulation (EU) No 10/2011.

The Matrix Project derived country data sets for Germany, France, Italy, Spain and United Kingdom with the respective packaging surface to which consumers are exposed per plastic material group and per consumed food and the respective calculation of LOIs.

Plastics material groups can be assessed on a country base to define the level where identified migrants need to be further risk assessed or not.

If NIAS assessments are addressed using the Matrix method the data and assessments become part of the supporting documentation of the products investigated at the respective stage in the Plastics value chain.

In general, the same methodology applied here for NIAS can be used for any non-listed substances.

B2. FACET Project

Within the 7th Framework Research Programme, Europe has developed a new tool for exposure of substances migrating from food contact packaging. FACET (Flavours, Additives and food Contact material Exposure Task) is an EU-funded project aimed at estimating exposure to three types of food chemicals: food additives, flavourings and migratable substances from food contact materials. The FACET project which was officially finished in October 2012 developed a software tool that models exposure to substances migrating from food contact material on a country base for the EU population. The probabilistic exposure results are based on comprehensive pan-European food consumption and food packaging data encrypted into the software.

See http://ihcp.jrc.ec.europa.eu/our_activities/food-cons-prod/chemicals_in_food/FACET

$$\text{EXPOSURE}_{\text{from FCMs}} = \text{MIGRATION (Conc. in Food)} \times \text{Food Consumption} \times \text{Packaging Material Usage}$$

Objective of WP4.2 in FACET project:
Development of a probabilistic model to establish concentrations of migrants from packaging materials in foods

2.3 Toxicological Assessment

Toxicological assessment aims to identify the adverse toxicological effects that a substance could cause (hazard Identification) and secondly, to define the critical dose or exposure level of a substance in the daily diet, below which the substance is not expected to pose a risk to human health (dose response assessment or hazard characterisation).

Most adverse effects for chemicals occur at a particular dose (Paracelsus: “dose makes the poison”). Toxicological studies or alternative data will be applied to derive the daily dose which can, based on conservative assumptions, be assumed with reasonable certainty to be safe.

This critical dietary exposure level is often referred to as the Tolerable Daily Intake (TDI), generally used for substances appearing in food but not intentionally added or the Acceptable Daily Intake (ADI) for substances intentionally added to food, usually expressed in mg/person/day or mg/kg bodyweight/day.

Based on the TDI and the European default assumption that a 60 kg person consumes a kilogram of food per day a self-derived Specific Migration Limit (SML) for the substance can be calculated using the following formula:

$$\text{Self-derived SML (mg/kg food)} = 60 \text{ (kg body weight)} * \text{TDI (mg/kg body weight/day)} / 1 \text{ kg food/day}$$

However, genotoxic mutagens and carcinogens are an exemption to this basic principle. For their mode of action, it is traditionally assumed that already one interaction event between a substance molecule and a DNA molecule could theoretically lead to an adverse effect, so that a no-threshold-mechanism is assumed⁴. Generally, the aim is to strictly avoid the presence of genotoxic mutagens and carcinogens in food contact materials. However, this may not always be possible, especially for NIAS. A safety assessment for such cases would follow the internationally accepted scientific principles of linear low dose extrapolation, the Margin of Exposure⁹ approach or Derived Minimal Effect Levels¹⁰ approach. In the case of food contact materials applied in Europe, the MOE approach is preferable, as it has been reviewed and recommended by the EFSA Scientific Committee¹¹.

Regulatory agencies in the United States, the Food and Drug Administration (FDA) and the European Union (EU) use a tiered approach based on the “dose makes the poison” principle to regulate substances that e.g. migrate from food packaging and processing equipment to food. Toxicological data may not be required when the exposure is extremely low. Under U.S. FDA guidance, substances with an exposure below the Threshold of Regulation of 1.5 µg/person/day and no concern of genotoxicity, do not require specific toxicological data. For Europe, under the provisions of the Regulation, substances that have not been evaluated and authorized and are not classified as carcinogenic, mutagenic or reprotoxic, can be used in plastics layers behind a functional barrier if they don't migrate at a detection limit of 10 µg/kg food. This “no-migration” concept for non-CMR substances has been adopted under the CEPE Code of Practice for non-listed substances in direct food contact coatings.

* There is on-going scientific debate about this hypothesis and the consensus may change in the near future, but this has to be discussed elsewhere and the current guidance document will build on the traditional hypothesis and risk assessment methods.

The first step of a safety assessment is always the search for toxicity data on the substance. Subsequently, there are basically two approaches to determine the dietary exposure thresholds for substances:

- The determination of a tolerable daily intake (TDI), based on toxicological studies performed on the substance or a structurally similar substance (read across) or
- if no substance specific data are available, use the *Threshold of Toxicological Concern* (TTC) concept as a basis

Substances being suspected or known genotoxins and/or carcinogens require specific risk assessment methodology which shall not be discussed here. For guidance, please refer to the MOE approach^{10, 11}.

2.3.1 Determination Tolerable Daily Intake (TDI) based on specific toxicological studies

Once it has been demonstrated that a substance does not pose any concern with regard to genotoxicity, an appropriate dose descriptor from repeated dose (chronic/subchronic/sub-acute) toxicological studies can be selected. Guidance on dose descriptor selection is for example available by the ECHA guidance on information requirements and chemical safety assessment Chapter R8.2, the ECETOC report TR 85 - Recognition of, and Differentiation between, Adverse and Non-adverse Effects in Toxicology Studies¹², ECETOC report TR 99 - Toxicological Modes of Action: Relevance for Human Risk Assessment¹³.

The Tolerable Daily Intake number can be derived from e.g. the NOAEL (No-Observed-Adverse-Effect-Level) or a benchmark dose (BMD-L) obtained from repeated dose (chronic/sub chronic/sub-acute) toxicological studies and taking into account certain assessment factors.

$$\text{TDI (mg/kg body weight/day)} = \text{NOAEL (mg/kg body weight/day)} / \text{assessment factor}$$

For EU food contact materials, the typical convention is to calculate the TDI by dividing the NOAEL obtained from an oral sub chronic (90 days) study with a default assessment factor of 100. This factor gives an additional margin to take into account the possibility that humans may be more sensitive than animals and that some humans may be more sensitive than others. The factor 100 is constituted of two factors of 10. One factor of 10 is intended to account for interspecies differences. This factor of 10 is envisaged as converting the findings in animals to equivalent findings in humans. A second factor of 10 is used to account for differences in typical humans and sensitive sub populations such as children, the elderly or compromised individuals.

These two assessment factors are intended to be conservative and address a wide range of chemicals. Recent guidance provided by the International Program on Chemical Safety (IPCS)¹⁴ and ECHA¹⁰ allows for deviation from the values of 10 when the data on the specific substance is sufficient to justify alternative values. In certain instances, smaller values can be justified using data on mechanism of actions or modeling of the pharmacokinetics of the compound.

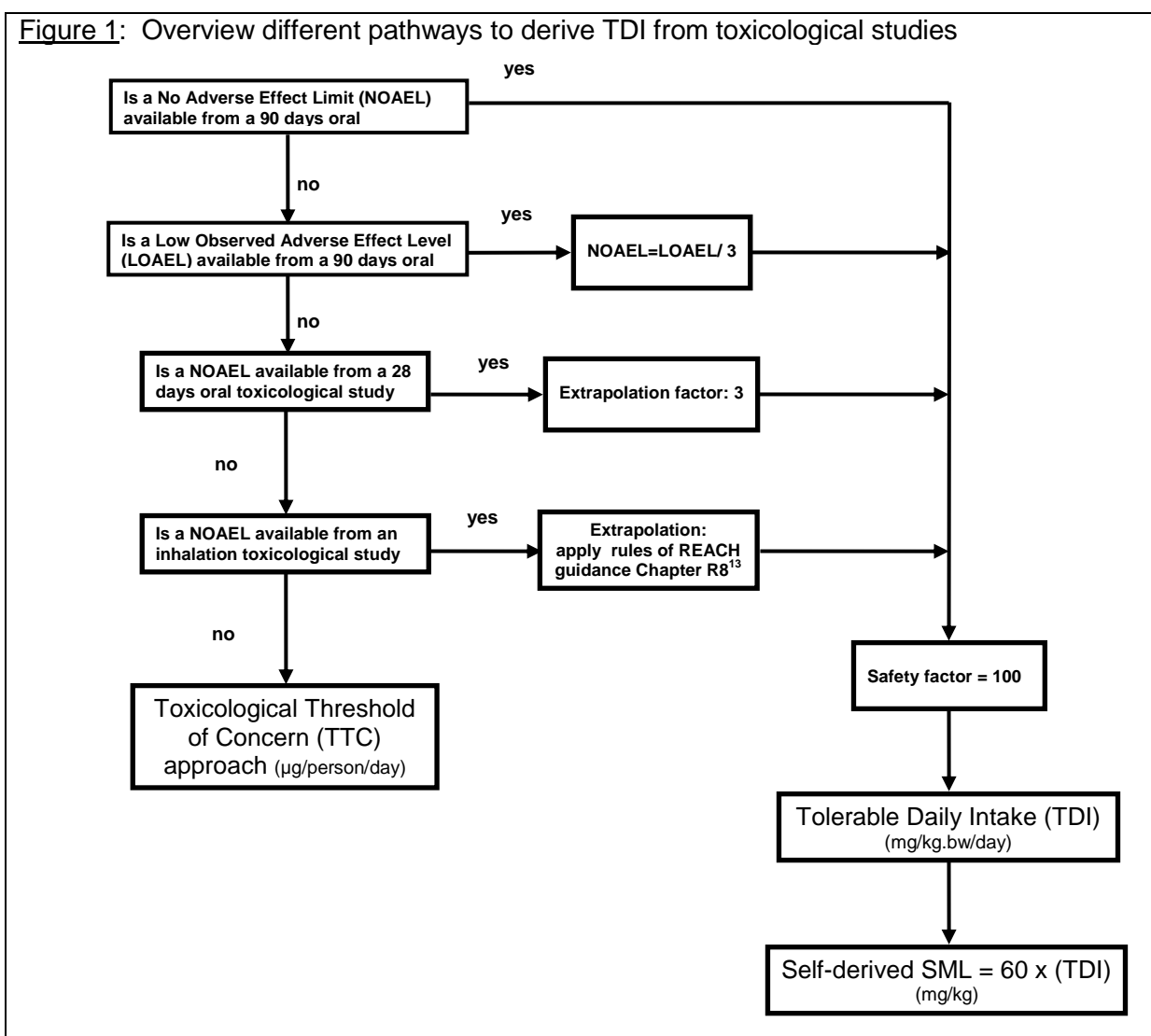
Additional assessment factors might be required in certain cases. Guidance on the need for additional uncertainty factors can be taken from the ECHA guidance for setting DNELs (Derived No-Effect Levels)¹⁰. Examples where additional factors may be required include the following cases:

- To reflect uncertainties/data gaps in the database (e.g. study not performed to current scientific standards, read-across study),
- Where a LOAEL (Lowest-Observed-Adverse-Effect-Level) is available instead of a NOAEL,

- When sub-acute (28 days) study is available and is used to estimate the NOAEL in a (sub)chronic study.

In case that oral study is not available, it may be possible to perform a route to route extrapolation¹⁰ to derive an oral NOAEL from a systemic NOAEL obtained from a non-oral exposure route (e.g. inhalation or dermal study).

Figure 1 provides an overview of various options to derive a TDI for a substance and the assessment factors to be applied. The Threshold of Toxicological Concern is discussed in the following section.



If no data on the substance to be assessed exist but data are available on a substance being structurally very similar, and based on a scientific rationale¹⁵ it can be shown with reasonable certainty that the toxicological properties of both substances are comparable, a NOAEL from the surrogate substance may be used to define the TDI of a substance. This process is referred to as a “read across”.

An additional source of toxicity data on substances or potential surrogates is the U.S. FDA webpage. This web page includes a database of Cumulative Estimated Daily Intakes (CEDIs) and Acceptable

Daily Intakes (ADIs) for a large number of food contact substances⁷. While the information can be useful, it has to be evaluated against the most recent toxicology studies on the substance of interest.

For substances for which toxicity data are available, it is important to use all the data for selecting the most appropriate NOAEL to determine the TDI.

It should be noted that the data used to derive the TDI for the calculation of the self-derived SML can be limited and lead to a relatively high TDI. Consequently on the basis of the set of toxicological data available, it may be needed to limit the TDI value and the derived SML according to the rules given e.g. by the EFSA Note of Guidance⁵

On the other hand, if a high TDI is needed, additional information / data on additional endpoints such as reprotoxicity studies, long term studies, etc. has to be considered.

2.3.2 Determination Threshold of Toxicological Concern (TTC)

The TTC is a risk assessment tool that, establishes human exposure levels for chemicals below which there is no appreciable risk to human health. It is a useful tool for assessing substances of unknown toxicity present at low levels in the diet where the structure of the compound^{16,17,18,19} is known, but no substance specific toxicity data or data on a similar substance exist.

The TTC approach is used by EFSA for the safety assessment of flavoring substances and metabolites, degradation and reaction products of pesticides as further applications like risk assessments for cosmetic ingredients, household products and impurities in therapeutic drugs. EFSA has established a TTC Working Group to study the applicability of the TTC approach for safety assessments for other applications including food contact materials²⁰. A task force at ILSI²¹ is working on further developing the science of the TTC concept. This approach was used first by the US FDA to establish the threshold of regulation for substances that migrate at 0.5 µg/kg or less into food wherein such substances are exempt from food additive regulations and procedures.

The TTC methodology is based on a decision tree approach (figure 2) that uses information on the molecular structure of a substance to assign the substance to one of a number of classes. Certain substance classes are exempt from the TTC concept, either because they were not part of the toxicological database used to derive the thresholds, or because they are of high toxicological concern and warrant a safety assessment based on substance specific data. The lowest thresholds for substances to which the concept can be applied are applied to a class of substances with functional groups that are structural alerts for genotoxicity. Substances without these alerts are assigned into one of several classes that are based on data from non-cancer endpoints in toxicological studies. These non-cancer classifications include the class of organophosphates and three broad classes of chemicals referred to as Cramer classes I, II, and III.

For illustration, table 1 lists the TTC thresholds as described by Kroes et al. 2004. The reader is directed to the original references of Kroes et al. 2004¹⁶ and Cramer and Ford 1978¹⁸ for additional information on using the concept. IT tools, e.g. ToxTree²² are available to facilitate the determination of the Cramer Class of a chemical.

See introduction and link to the ToxTree free software on reference 22.

⁵ <http://www.efsa.europa.eu/en/search/doc/21r.pdf>

Table1: TTC exposure thresholds as described by Kroes et al. 2004¹⁶. Each threshold requires the classes listed below its level, to be excluded.

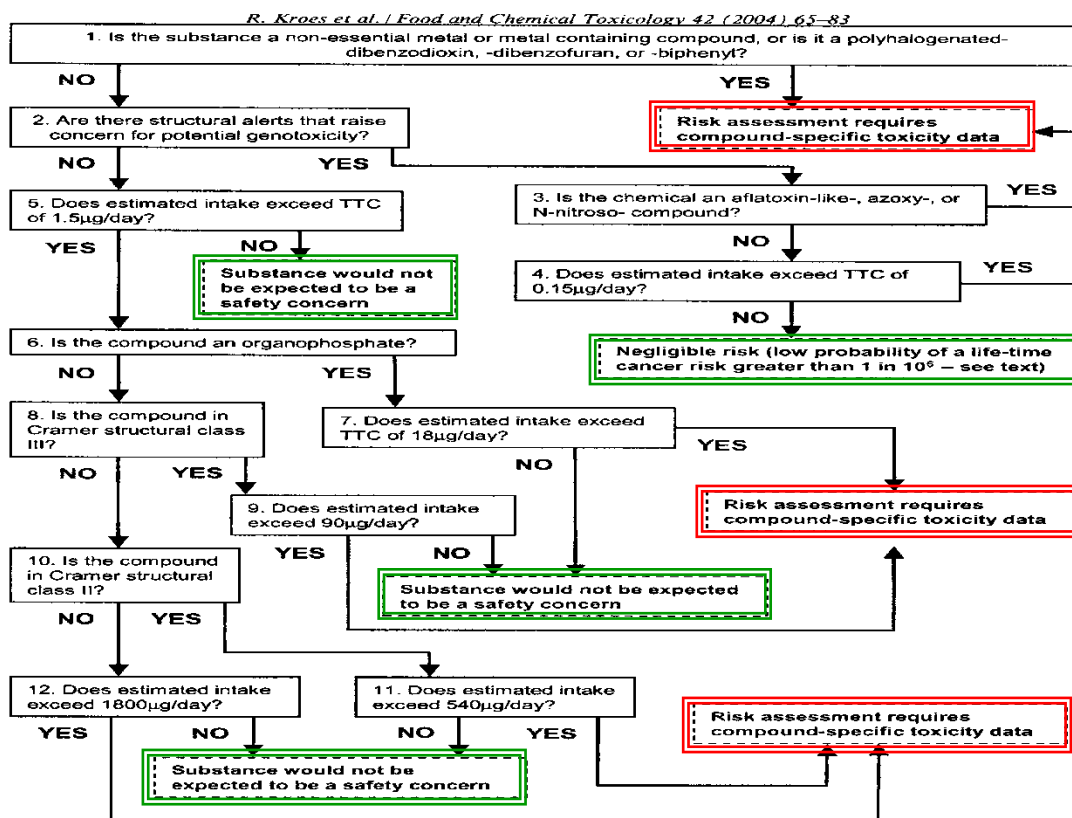
Structural class	Description	TTC exposure limit (µg/person/day)
Cramer class I (least toxic)	Substances with simple chemical structure and for which efficient modes of metabolism exist, suggesting a low order of oral toxicity.	1800
Cramer class II (intermediate)	Substances which possess structures that are less innocuous than class I substances, but do not contain structural features suggestive of toxicity like those substances in class III.	540
Cramer class III (most toxic)	Substances with chemical structures that permit no strong initial presumption of safety or may even suggest significant toxicity or have reactive functional groups.	90
Organophosphates	Organophosphate structures which may have neurotoxic properties	18
Threshold of regulation	Substances for which there are no structural alerts for genotoxicity	1.5
Genotoxicity alerts	Substances for which there are structural alerts for genotoxicity but which are not aflatoxin-like, azoxy- or N-nitroso-compounds	0.15

Substances for which there are no structural alerts for genotoxicity and which do not belong to one of the TTC exempt classes (e.g. non-essential metals) are considered safe if the exposure is below 1.5 µg/person/day. The US FDA regulation has specified this exposure limit in the law (CFR 21 par 170.39) as the so-called “Threshold of Regulation” sometimes also referred to as “Threshold of Regulatory Concern” (TORC). For substances with structural alerts for genotoxicity an even lower exposure threshold of 0.15 µg/person/day has been defined.

Please note that the tables are not synchronized with the EFSA opinion on TTC (see further for the reference to that opinion). The EFSA opinion suggests that Cramer Class 2 intermediates should be in Cramer Class 3.

Currently there is no specific guidance from DG Sanco so it is suggested that any suitable method of risk assessment be used (EU or US)

Figure2: Decision tree TTC¹⁶



A special challenge may occur when several non-intentionally added substances migrate at very low level. First, albeit difficult from analytical perspective, the assessment requires certain knowledge on the substances structures to be able to search for toxicity data or to apply the TTC concept¹. With respect to the latter, the assessor would have to have sufficient structural information/knowledge on the origin of the substances to answer the questions of the TTC decision tree.

2.4 Risk Characterisation

In the final risk characterisation step, the typical exposure level to the substance in the daily diet (the Estimated Daily Intake) is compared to the maximum tolerable exposure level (the Tolerable Daily Intake). Alternatively the migration level into the food is compared to a self-derived Specific Migration Limit (SML). As long as the Estimated Daily Intake is below the Tolerable Daily Intake or the migration level under the typical condition of use is below the self-derived SML, the use of the substance is considered safe.

$$EDI < TDI \quad \text{or} \quad \text{Specific migration} < \text{self-derived SML} \Rightarrow \text{PRODUCT IS SAFE!}$$

In case the product is not considered safe in the application, then either the risk assessment needs to be refined (refine the exposure estimation or generate more toxicological data) or the exposure to the substance (migration level) needs to be reduced.

The risk assessment must be reviewed regularly to take into account the evolution of the knowledge relating to the toxicity of the substance and if the conditions of use are changed or different.

2.5 Reporting of risk assessment in supporting documents to Declaration of Compliance (DoC)

The risk assessments on non-listed substances should be reported in the Supporting Documentation to the Declaration of Compliance (DoC) according to Article 16 of the Regulation. The format and content of the Supporting Documentation should be defined on a case-by-case basis. Below are some suggestions about information to be included as applicable:

- Identity and the address of the company
- Date of the risk assessment
- Authors
- Identity of the substance:
 - Trade name
 - Chemical name
 - CAS number
 - Chemical structure
 - Molecular weight
 - Purity criteria and the main impurities
- Physical and chemical properties available (especially partition coefficient or solubility, boiling point...)
- Thermal stability and/or decomposition products (for example for organic peroxide) if available and appropriate
- Conditions of use in the polymer:
 - Description of the conditions of use
 - Function of the substance in the manufacturing process and in the final polymer
 - Maximum concentration in the final material
 - Specific steps used to minimize the residual amount of the substance in the final polymer...
- Toxicological data:
 - Cramer class and TTC exposure limit if TTC approach is used
 - Read across if appropriate and arguments used to do this read across
 - Toxicological data if available:
 - Genotoxicity and mutagenicity
 - NOAEL
 - TDI derived
 - SML calculated based on toxicological data
- Evaluation for the Exposure:
 - Method used to determine the migration level:
 - Worst-case calculation
 - Migration modeling
 - Migration tests: conditions used (time, temperature, nature of the simulants)
- Conclusions
- List of references (if appropriate)

3 Conclusion

The European food contact legislation is a risk-based legislation, which means that risk assessments have to be carried out to demonstrate safety.

Under the Framework Regulation (EC) No 1935/2004 all food contact materials have to be risk assessed in order to demonstrate safety. Commission Regulation (EU) No 10/2011 on plastics materials for food contact has a regularly updated list of approved monomers and additives whilst Article 19 of this Regulation outlines that certain categories of non-listed substances and so-called NIAS (non-intentionally added substances) have to be risk-assessed by industry in accordance with internationally recognized scientific principles. This risk assessment on these substances is needed as part of the supporting documentation to the Declaration of Compliance (DoC) for the plastic material to be shown to authorities upon their request.

A risk assessment generally consists of four components: literature survey on existing legislation, exposure assessment, toxicological assessment and risk characterisation.

Like for listed substances, exposure assessments for non-listed substances require information on migration of these substances into food and data on daily exposures (food consumption and packaging data). There are various ways of calculating exposure: US FDA, industry exposure tools (so-called Matrix Tool), EU Facet tool.

Toxicological assessments will be applied to derive the daily dose which can be assumed to be safe. The first step of such a toxicological safety assessment is the search for or the determination of toxicity data on the substance. There are basically two approaches to determine the dietary exposure thresholds for substances: (1) based on toxicological studies performed on the substance or a structurally similar substance (read across) and (2) based on the threshold of toxicological concern (TTC) concept if no substance-specific data are available. The threshold of toxicological concern is determined based on the chemical structure known of the substances using dedicated free software.

In the final risk characterisation step, the typical exposure level to the substance in the daily diet (the Estimated Daily Intake) is compared to the maximum tolerable exposure level (the Tolerable Daily Intake). Alternatively the migration level into the food, calculated (worst-case or modelling) or measured is compared to a self-derived Specific Migration Limit (SML) applying specific rules. As long as the Estimated Daily Intake is below the Tolerable Daily Intake or the migration level under the typical condition of use is below the self-derived SML, the use of the substance in the defined plastics materials taking into account the defined use conditions is considered safe for consumer health.

4 Annex (examples)

Note: Solely for the purpose of illustration, the examples provided below are assessed both via substance specific data and the TTC approach. While this demonstrates the conservatism inherent to the TTC approach, it must be clearly stated that substance-specific data are always the first choice. The examples given below are based on real cases that have been presented as generically as possible for avoiding to divulgate proprietary information.

Example 1:

Risk assessment on the use of a solvent as processing aid in a plastic food contact material

Liquid paraffin solvents, known also under the designation of linear or branched aliphatic hydrocarbons or alkanes

, could be used as desensitizing agent and/or carrier of organic peroxides in the manufacturing process of several polymers like polyolefins, PVC, ethylene copolymers, rubbers etc., intended for food contact applications. The use of desensitizing agents is essential to comply with the transportation regulation relating to dangerous substances and for safety reasons. The boiling point of the aliphatic hydrocarbons used is often above 150° C following the requirements of the transportation regulation of dangerous chemicals. Due to the high boiling point of these aliphatic hydrocarbons, they could be present in the final plastic materials intended for food contact applications. Alkanes don't provide physical or chemical properties to the final materials and are considered as Aids to Polymerisation exempted from positive listing under Regulation (EU) No. 10/2011. These solvents present as impurity in the manufacturing process, have to be risk assessed according to the Article 19 of the Regulation to guarantee the safety of the final material.

The Cramer Class of liquid alkanes is I. Consequently the exposure threshold for aliphatic hydrocarbons is 1800 µg/person/day or 1.8 mg/person/day. So, for a European eating 1 kg food per day, the SML based on TTC approach will be:

$$\text{SML}_{\text{TTC}} = 1.8 \text{ mg/kg food}$$

For one of the aliphatic hydrocarbons currently used in Food Contact Materials, a NOAEL value has been determined of 330 mg/kg bw/day, based on a 90 days sub chronic oral toxicity study on rats. This alkane is considered as non-genotoxic (data available). Following the pathway given by the Figure 1, the Tolerable Daily Intake can be derived by dividing the NOAEL by the Safety Factor 100.

$$\text{TDI} = \text{NOAEL} / 100 = 3.3 \text{ mg/kg bw/day}$$

The self-derived SML, for a 60 kg weight European eating 1 kg of food per day will be:

$$\text{Self-derived SML} = \text{TDI} * 60 = 198 \text{ mg/kg food}$$

The comparison of these two SML's shows that the TTC approach is conservative compared to real toxicological data. The SML calculated, on the base of toxicological data, is higher than the overall migration limit of 60 mg/kg food set by the Regulation.

The maximum concentration in the plastic of aliphatic hydrocarbons coming from organic peroxides used during the different manufacturing processes is less than 3500 mg/kg. But the synthesis is often followed by a finishing step inducing a reduction of the residual concentration of aliphatic hydrocarbons at a level below 1000 mg/kg.

Assuming 100 % migration from a 6 dm² packaging having 100 µm thickness and a density between 1 g/cm³ to 1.4 g/cm³, the migration level according to the formula in section 1.1.1 will be **between 6 and 8.4 mg/kg food**. This worst-case migration is a factor 23 to 33 below the SML of 198 mg/kg food as derived from toxicological data and above the SML_{TTC} of 1.8 mg/kg food given by the TTC approach.

As the TTC approach is very conservative compared to the real toxicological data measured on the substance, the conclusion of the risk assessment based on real tox data is that final materials containing the assessed aliphatic hydrocarbons are safe for the consumer in the intended uses.

Example 1.2

Packaging: High impact polystyrene container

Food: yoghurt

Ethylbenzene (residual solvent)

Initial level of in yoghurt pot: 170 ppm

Pot dimensions:

- Maximum diameter: 4,5 cm
- Height: 6,3 cm
- Surface volume ratio: 1,05 cm⁻¹
- Bottle thickness: 0,02 cm

Contact temperature: between 5°C and 20°C

Contact time between 0" and 1 month

Toxicity assessment: (Toxtree)

Cramer class II. Class III taken as a conservative approach: 1,5 µg/kg.bw/day.

Genotoxic: no

Carcinogenic: no

Toxic to reproduction: no

Teratogenic: no

Toxic to development: no

Exposure estimated using FACET software (version 2.0.5)

Total population: 1,145 µg/kg.bw/day

Consumers: 1,643 µg/kg.bw/day

Conclusion:

Whereas the average exposure of the total population is below the Cramer class III threshold, the average exposure of consumers is above. In these conditions the risk assessment shall be refined to prove that in the conditions detailed above, the final material is safe for the consumers. If the material is not always safe, the manufacturer of the polymer will have to reduce the level of solvent used to drop the level of the degradation product below the threshold.

Example 2:

Risk assessment on a surfactant used in the manufacturing of a polymeric additive used in a plastic food contact material

Impact modifiers used as a polymeric additive in different Food Contact Materials are often manufactured using an emulsion, suspension or micro suspension polymerisation process. Surfactants or generally surfactant systems are used to disperse the monomers into the water. According to the Regulation (EU) No 10/2011, the surfactants can be considered Polymer Production Aids in these manufacturing processes. Certain surfactants are used as additives in food contact plastics and are listed in Annex I of the Regulation. Others are not listed and require a risk assessment according to Article 19 of the Regulation.

Impact Modifiers are often acrylic and methacrylic copolymers with styrene, butadiene and other specific monomers such as cross-linking agents. Depending on the manufacturing process, the concentration of the surfactants system may vary between 1 to 10 % of the total active ingredients of the recipe excluding water that is eliminated during the finishing and drying steps.

For this example, based on a real situation, the maximum concentration of one proprietary surfactant, already used in cosmetic formulations and in certain Food Contact Materials, is below 1.5% in the final pure acrylic-methacrylic copolymer used as impact modifier. In the normal conditions of use, the maximum level of the impact modifier into the final plastic material polymer is below 5%.

The Cramer Class of the specific surfactant used, on the basis of its chemical structure, is III. Consequently the exposure threshold for the substance is 90 µg/person/day or 0.09 mg/person/day. So, for a European eating 1 kg food per day, the SML based on TTC approach will be:

$$\text{SML}_{\text{TTC}} = 0.09 \text{ mg/kg food.}$$

For this proprietary surfactant a NOAEL value of 155 mg/kg bw/day has been determined derived from a 90 days sub chronic oral toxicity study on rats. Based on 3 negative mutagenicity studies, this surfactant is considered as a non-genotoxic. Following the pathway given by the Figure 1, the Tolerable Daily Intake can be derived from the NOAEL by applying the safety factor 100.

$$\text{TDI} = \text{NOAEL} / 100 = 1.55 \text{ mg/kg bw/day}$$

The self-derived SML for a 60 kg European eating 1 kg of food per day will be:

$$\text{Self-derived SML} = \text{TDI} * 60 = 93 \text{ mg/kg food}$$

The comparison of these two SML's shows again that the TTC approach is conservative compared to real toxicological data. The SML calculated, on the basis of toxicological data is higher than the overall migration limit of 60 mg/kg food set by the Regulation. As the full set of toxicological data is not available according to EFSA rules for this risk assessment, the SML will be limited to 5 mg/kg.

The concentration of this specific surfactant used to prepare an acrylic-methacrylic copolymer is below 2 % of the active ingredients (monomers, additives, PPA's, AP's excluding water). The commercial product is a water solution containing 30 % the surfactant. The maximum level of the surfactant in the acrylic-methacrylic copolymer is 0.6%. After the different finishing steps, this concentration could be lower. The impact modifier is often incorporated at 2 % in a PVC resin. So in the PVC polymer the maximum concentration of the surfactant is: **120 mg/kg (2% *30% * 2%)**.

Assuming a 100 % migration from a 6 dm² surface packaging of 100 µm thickness and 1.4 g/cm³ density, the migration level into the food will be: **1 mg/kg food**.

This worst-case migration is a factor 5 below the limited SML as derived from toxicological data and far above the SML_{TTC} of 0.09 mg/kg food given by the TTC approach. Even if the safety factor may be considered as lower, because it is based on a worst-case scenario which over-estimates the migration, the safety of the final material has been demonstrated by modeling that is not presented in this document

As the TTC approach is very conservative compared to the real toxicological data measured on the substance, the conclusion of the risk assessment based on real tox data is that the final materials containing this specific surfactant as component in the impact modifier are safe for the consumer in the intended uses.

Example 3: Examples of risk assessment of degradation products from antioxidants

Antioxidants are commonly added to polymers to first protect them against thermal degradation during their processing and also to reduce further degradation during their shelf-life. Phenolic antioxidants belong to the second class. The two following examples illustrate the risk assessment procedure which can be followed to demonstrate the absence of safety concern regarding trace levels in a food packaging of two typical degradation products from phenolic antioxidants.

Example 3.1:

Packaging: High Density Polyethylene (HDPE) 1L bottle

Food: orange juice

Degradation product: 2,6-di-tert-butylphenol (CAS n°:128-39-2)

Mw: 206.32 g/mole

Log p = 0

Initial level of the degradation product in the bottle: 10 ppm

Bottle dimensions:

- Maximum diameter: 8.4 cm
- Height: 18.04 cm
- Surface volume ratio: 0.531 cm⁻¹
- Bottle thickness: 0.04 cm

Contact temperature: between 20°C and 40°C

Contact time between 0" and 6 months

Toxicity assessment: (Source: ECHA)

Oral toxicity: NOEL = 100 mg/kg.bw/day derived TDI = 1 mg/kg.bw/day (Safety factor: 100).

Genotoxic: no

Carcinogenic: no

Toxic to reproduction: no

Teratogenic: no

Toxic to development: no

Exposure estimated using FACET software (version 2.0.5)

Total population: 0,0001562 mg/kg.bw/day

Consumers: 0,0003166 mg/kg.bw/day

Safety margin:

Total population: 1/0,0001562 = 6402

Consumers: $1/0.0003166 = 3158$

Conclusion:

Sufficient margin of safety has been demonstrated (exposure is far below the TDI). So the food packaging containing this decomposition product of an antioxidant is safe for the consumers in the intended use detailed above.

Example 3.2:

Packaging: High impact polystyrene container

Food: yoghurt

Degradation product: 3,5-di-tert-butyl-4-hydroxyacetophenone (CAS n°:14035-33-7)

Mw: 248.37 g/mole

Log p = 0

Initial level of the degradation product in the pot: 15 ppm

- Pot dimensions:
- Maximum diameter: 4,5 cm
- Height: 6,3 cm
- Surface volume ratio: $1,05 \text{ cm}^{-1}$
- Pot thickness: 0,02 cm

Contact temperature: between 5°C and 20°C

Contact time between 0" and 1 month

Toxicity assessment: (Toxtree)

Cramer class II → Class III taken as a conservative approach (EFSA opinion): 1.5 µg/kg.bw/day.

Genotoxic: no

Carcinogenic: no

Toxic to reproduction: no

Teratogenic: no

Toxic to development: no

Exposure estimated using FACET software (version 2.0.5)

Total population: 0.101 µg/kg.bw/day

Consumers: 0.145 µg/kg.bw/day

Conclusion:

No consumer safety concern as the exposure is below the Cramer class III threshold. So the food packaging containing this decomposition product of an antioxidant is safe for the consumers in the intended use detailed above.

Example 3.3:

Note: the input data chosen in the above three examples do not necessarily represent the actual commercial products.

Example 4: Ziegler-Natta and metallocene catalyst systems used in food contact packaging plastics - How consumer safety is best ensured?

Most legislative frameworks on food contact materials rely on safety evaluation of the main constituents by authorities. Minute amounts of minor constituents, for example catalyst residues, are typically not included in mandatory listing schemes. With the growing concern for safety, it is

often suggested that extending positive listing to catalysts would enhance consumer safety. This paper takes the example of Ziegler-Natta catalysts commonly used in polymerisation of polyolefins and suggests that proper safety evaluation needs to rely on the complete catalyst system details in combination with the production process parameters.

Food contact material safety principles:

In legislative frameworks for food packaging materials, safety is ensured by two general principles: inertness and safety of the material in use. In very simple terms, this means that materials should not release their constituents in amounts that could endanger human health.

For monomers and additives used in food contact plastics these principles are translated into positive listing of substances in Annex I of EU Regulation No 10/2011. After a review of the hazard level of substances, risk management measures are prescribed, typically in the form of Specific Migration Limits, expressed as mg/kg food.

For other substances, such as aids to polymerisation, positive listing systems at EU or at national level may not provide the desired safety objective. The example of Ziegler Natta catalysts that are used in many polyolefin production systems is presented and the importance of evaluating the whole catalyst system in conjunction with production conditions is highlighted.

Aids to Polymerisation, Ziegler-Natta systems description:

Ziegler-Natta catalyst systems are widely used in the polymerisation of polymers of α -olefins. These systems rely on metal chlorides or other metal complex compounds with a co-catalyst to provide highly efficient and selective polymerisation characteristics to generate well controlled polymeric structures. Complete catalyst systems can consist of several components:

- Catalyst itself can be: TiCl_3 , TiCl_4 , VOCl_3 , VOCl_4 , Zr or Hf. Metallocene catalysts use a variety of complexes of metals, ranging from scandium to lanthanoid and actinoid metals, and a large variety of ligands.
- Catalysts are usually used in conjunction with co-catalysts, typically organoaluminum compounds.
- Catalyst supports are also used and are usually MgCl_2 .
- Most catalyst heterogeneous systems include a carrier, a material that determines the size of catalyst particles. Carrier can be micro porous spheres of amorphous silica.
- Another component can be organic modifiers, usually an ester of an aromatic di-acid or a di-ether.
- The complete system can also include solvents, pH buffers, acid scavengers, anti-static compounds ...

Ziegler-Natta catalyst systems are complex with a very high number of possible combinations and proportions of components. Additionally, most catalysts and organo aluminum co-catalysts are unstable in air and pyrophoric and the catalysts need to be prepared and handled under an inert atmosphere. Catalyst systems do not survive as such after polymerisation and produce potential NIAS from decomposition and reaction products. These reaction products are not only function of the catalyst but can also be influenced by the particular conditions used in the production process. Consequently an evaluation based on starting substances can be misleading and dismiss the basic objective of safety evaluation of migrants.

Safety assessment approach:

Given the complex nature of catalyst systems and the importance of process parameters, it is important to base safety assessment on residual reaction products. These products are often not available in isolation and traditional toxicological evaluation might be challenging or even impossible.

However, safety assessment based on internationally recognized principles according to Article 19 of EU Regulation 10/2011 provides adequate evaluation basis.

For example, high efficiency of metallocene catalysts (more than 300 000 kg polymer / kg catalyst) can be assessed by using Threshold of Toxicological Concern (TTC) methodology after proper verification that exclusion criteria are satisfied (absence of genotoxicity, organophosphates and carbamates ... in the reaction products).

Conclusion

A safety assessment scheme for catalysts based on listing of starting substances would generate an impractical and burdensome evaluation system. More importantly, it is likely to fail to capture all the relevant catalyst system combinations and their particular process conditions. Evaluation of reaction products stemming from actual catalyst systems and production parameters is far more effective particularly when using internationally recognized principles according to Article 9 of EU Regulation No 10/2011.

Example 5: Organic peroxides used for initiating radical polymerisation to prepare polymers intended to come into contact with food : how to risk assess organic peroxides and their decomposition to comply with article 3?

Introduction

According to the framework regulation (EC) No 1935/2004 on food contact materials, the safety evaluation of the main constituents is carried out by European Food Safety Authority (EFSA). Minute amounts of constituents, for example, organic peroxides, are used as Aids to Polymerisation and are typically not included in mandatory listing schemes. To improve consumer safety, it is often suggested that DG Sanco extend positive listing to include Aids to Polymerisation. The objective of this example organic peroxides, are commonly used to initiate radical polymerisation of some polymers used for food packaging. The example will demonstrate that a proper safety evaluation needs to rely on the real production process parameters of the polymer in combination with the properties of the Aid to Polymerisation used.

Food contact material safety principles:

The Inertness and the safety of the materials intended to come into contact with foodstuffs are the two general principles established in the legislative framework for ensuring the consumer safety. This means that food contact materials, but also drinking water materials, should not release their constituents in quantity that could endanger the consumer health and modify the composition and the organoleptic properties of the foodstuffs.

In the Regulation (EC) No 10/2011 that is a specific measure for one specific class of food contact materials; these two general principles have been transposed into a single European list of monomers and additives. After a review of hazard of each substance of these two families of constituents, risk management measures are prescribed currently in the form of Specific Migration Limits generally expressed in mg/kg food, with or without restrictions and specifications, on the basis of the technical information in the authorisation dossier submitted by industry to EFSA.

For other substances, such as aids to polymerisation, a systematic positive listing approach at EU or at national level might not achieve the desired safety objectives. The particular example of Organic Peroxides used for initiating radical polymerisation or for cross linking many of food contact polymers, is presented. The importance of evaluating the initiating system in relation with the polymerisation conditions is explained.

Aids to Polymerisation, organic peroxides:

Organic peroxides (OPs) are organic compounds containing peroxide functional group (-O-O-). The peroxide functional group is relatively weak and the -O-O- bond easily breaks when it is heated. This functional group is thermally unstable. This thermal decomposition generates free radicals having the general structure RO \cdot .

The principle purpose of organic peroxide is to decompose and to generate radical chemical reactions. In doing so it generates useful radicals that can be used as:

- Initiator to start radical polymerisation of several types of polymers ((metha)acrylic, PVC, EVA copolymers, fluorinated polymers...),
- Modifier to graft unsaturated monomers on polymer backbone,
- Viscosity modifier (visbreaking agent) to reduce the viscosity of polyolefines by breaking the polymer chains,
- Cross linking agent for thermosetting polymers (e.g.: polyester resins, silicones),
- Vulcanizing agent for the production of elastomers (e.g.: ethylene-propylene copolymers or terpolymers)

There are used in several totally different manufacturing processes of polymers like:

- Mass, emulsion, suspension or high pressure polymerisation
- Grafting in solution
- Reactive extrusion

Potentially, 68 organic peroxides can be used in the radical polymerisation of several polymers intended to come into contact with foodstuffs. All these organic peroxides can be classified into 8 families based on the chemical structure:

- Dialkyl peroxides (R'-(OO-R)_x)
- Diacyl peroxides (R-CO-OO-CO-R')
- Hydroperoxides (R-OO-H)
- Ketone peroxides
- Peroxymonocarbonates (R-O-CO-OO-R')
- Peroxydicarbonates (R-O-CO-OO-OC-O-R')
- Perketal (R'-OO-R-OO-R')
- Peroxyesters (R'-CO-OO-R)

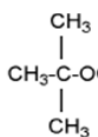
Decomposition products

Due to the chemical nature of this family of aids to polymerisation, organic peroxides generate decomposition products (free radicals) depending on the chemical structure of the OPs and on the production process parameters (time, temperature, pressure...). Some of them are small molecules with a low boiling point and may be easily eliminated during the different manufacturing processes of the polymers. Furthermore some of these decomposition products may be on the Union list of the regulation (EC) No 10/2011 (e.g.: 1-propanol, 2-ethyl-1-hexanol, acetaldehyde, acetic acid, acetone, benzoic acid, pentaerythritol, propylene...) and are approved for food contact plastics. The main thermal decomposition products are a part of the knowledge of the organic peroxides producers. This information may be found in the technical datasheet of the organic peroxides or may be delivered by the producers on request.

However a part of these free radicals may react with other constituents of the reaction medium and generate new species. These decomposition products are a part of the NIAS that may migrate into the food and shall be assessed according to the article 19 of the regulation (EC) No 10/2011 in accordance with internationally recognised scientific principles on risk assessment (see PlasticsEurope Guideline).

So only the company manufacturing the polymer and knowing the real conditions of uses, the recipe of the polymer and the production process parameters can do the risk assessment of the decomposition products generated by the organic peroxide in their conditions.

For example Di-ter-butyl peroxide (CAS: 110-05-4) belonging to the dialkyl peroxides family generates itself, by thermal decomposition (primary and secondary reactions), the following substances: Methane (CAS: 74-82-8), Acetone (on the Union positive list FCM: 119), T-butanol



(CAS: 75-65-0). The free radical obtained by β -scission of di-ter-butyl peroxide induces the production of an amount of T-butanol that is not on the Union list of the regulation (EC) No 10/2011. Assuming that the organic peroxide is decomposed only according this primary decomposition reaction (β -scission), it is easy to calculate the theoretical maximum concentration of the free radical: 1 mole of Di-ter-butyl peroxide generates 2 moles of T-butanol. This maximum concentration of the decomposition product may be used as a worst-case scenario for evaluating the consumer safety exposed to T-butanol assuming that the associated free radicals do not react with the monomers.

Half-life time of the organic peroxides

Another important characteristic of the organic peroxides is the half-life time ($t_{1/2}$). Half-life time is a convenient index representing the time at which 50 % of the organic peroxide has decomposed at any temperature. Half-life time is measured using a solution (generally of 0.1 mol/l and occasionally 0.05mol/l) of peroxide with a solvent relatively inert to radicals under nitrogen sealed in a glass ampoule, and immersed in a constant temperature bath set to the temperature required.

The decomposition of organic peroxide can be treated approximately as a first order reaction. The mathematical relation between the temperature of the manufacturing process and the half-life is the following:

$$k_d t = \ln(C_0/C_0-x)$$

k_d : decomposition rate constant

C_0 : initial concentration of organic peroxide

t : time

x : concentration of organic peroxide at t

The half-life time is the time for which $x = C_0/2$. Consequently the half-life time is:

$$k_d t_{1/2} = \ln(2)$$

The rate constant (k_d) is given as follows:

$$k_d = A e^{-E_a/RT}$$

A : frequency factor (s^{-1})

E_a : Activation energy (J/mole)

R : gaz constant (8.3142 J/mole.K)

T : temperature (Kelvin)

These parameters can be delivered upon request to the organic peroxides manufacturers.

Knowing the production process parameters (T , t), the half-life can be calculated at this temperature. The percentage of decomposed organic peroxide is totally correlated to the number of half-lives (see table 1). However, the polymerisation process is often followed by successive manufacturing processes (compounding process, extrusion or injection process of the final material...). These successive processes have an influence of the residual concentration of organic peroxide and decomposition products but may also generate other new NIAS.

Number of Half-life	Percentage of decomposed peroxide
1	50.00 %
2	75.00 %
3	87.50 %
4	93.75 %
5	96.90 %
6	98.60 %
7	99.20 %
8	99.60 %
9	99.80 %
10	99.90 %

Table 1: Number of Half-lives versus the percentage of decomposed peroxides

As example for the Di-2-ethylhexyl peroxydicarbonate (CAS: 16111-62-9), the activation energy and the frequency factor are the following and the half-life at different temperatures is given in the table 2.

- $A = 1.83 \cdot 10^{15} \text{ s}^{-1}$
- $E = 122.45 \text{ kJ/mole}$

T (°C)	T (°K)	$K_d \text{ (s}^{-1}\text{)}$	Half-life (s)
57	330.15	7.7E-05	8954
70	343.15	4.2E-04	1652
100	373.15	1.3E-02	52
130	403.15	2.5E-01	2.8
150	423.15	1.4E+00	0.5
180	453.15	1.4E+01	0.05
190	463.15	2.8E+01	0.024
200	473.15	5.5E+01	0.012

Table 2: Half-life of Di-2-ethylhexyl peroxydicarbonate at deferent temperatures

Di-2-ethylhexyl peroxydicarbonate may be used to initiate de radical polymerisation of PVC resins according to BfR II recommendations. The classical polymerisation conditions of the PVC manufactured with mass process are: 6 hours (21600 s) at 57° C followed by a post curing 15-20 minutes at 70° C.

At 57° C, the half-life of this organic peroxide is 8954 s. The polymerisation time represents 2.4 half-lives. At the end of the polymerisation step, 81.2 % of the initial concentration of the peroxide is decomposed and the polymer contains 18.8 % of the initial concentration of the peroxide. This polymerisation step is followed by the post curing step at 70° C, the half-life 27.5 minutes (1650 secondes). So, half of the residual concentration of peroxide is decomposed during the post curing step and the maximum residual concentration of organic peroxide in the base resin is 9.4 % of the initial concentration. This PVC resin is compounded with additives at 200° C for 20 minutes. At this temperature the half-life is 0.012 s. The compounding time represents 100,000 half-lives. In conclusion the organic peroxide used in the manufacturing process of PVC is totally decomposed after the compounding step and is not present in the final food contact articles. A similar approach can be used to risk assess organic peroxides used in the manufacturing process of other polymers.

Desensitising agents

By nature organic peroxide are unstable chemicals. The Union Nations have developed rules for the transportation and storage of organic peroxides and these have been transposed into Directive

2008/68/EC. To improve the thermal stability of peroxides, the use of desensitising agents is recommended by UN legislation. The desensitisers that may be used are the following:

- Inert organic solvents (e.g. mineral oils, some of them may have a boiling point above 150° C)
- Water
- Inert solids (e.g.: silica, calcium carbonate, polymers...)

If water is used as desensitising agent, a surfactant or suspension agent may be added according to the solubility of the organic peroxides into water.

Desensitising agents may be present in the polymer and must be also assessed.

Safety assessment approach:

In the normal condition of uses, the organic peroxides are totally decomposed and are not intended to be present in the final food contact materials. Several national recommendation or legislations introduce the legal requirement for verifying that the surface of the finished food contact materials shall not be tested positively for organic peroxide (see BfR Recommendations, Dutch Warenwet, Resolution of Council of Europe AP(92)2).

Most of the decomposition products react with monomers and the growing polymer chains during the polymerisation process and any un-reacted decomposition products will be eliminated in the successive processes up to production of the final food contact articles (compounding, manufacture of intermediate forms, manufacture of the final article). Decomposition products are not intended to be present in the food contact materials. Desensitizing agents may be present in the food contact materials and must also be also risk assessed (see example 1 hereinabove).

A safety assessment based on internationally recognized principles according to Article 19 of EU Regulation 10/2011 may be done by the polymer producers using real production process parameters with adequate information supplied by the organic peroxide producers.

Conclusion:

A safety assessment scheme for organic peroxides based added substances would generate an impractical and burdensome evaluation system because the only the organic peroxide itself would be assessed. Using internationally recognized principles, according to Article 19 of EU Regulation No 10/2011 for evaluation of decomposition products and reaction products stemming from initiating system and production parameters is a far more effective method of ensuring consumer safety.

5 Tools for risk assessment

OECD QSAR toolbox:

<http://www.oecd.org/chemicalsafety/risk-assessment/theoecdqsartoolbox.htm>

Joint Research Centre (JRC) – Guidance on migration modelling

Joint Research Centre (JRC) Migration testing guidelines

http://ihcp.jrc.ec.europa.eu/our_labs/eurl_food_c_m/guidance-documents

Commercial migration modelling institutes (FABES and AKTS)

<http://www.fabes-online.de/profil.php?lang=en>

<http://www.akts.com/>

Non-commercial modelling institute: INRA

<http://sfpp3.agroparistech.fr/>

Matrix data website

Not available yet.

Basic tox data: ECHA website:

<http://echa.europa.eu/>

Toxtree

Toxtree is a flexible and user-friendly open-source application that places chemicals into categories and predicts various kinds of toxic effect by applying decision tree approaches. The software is free. Details can be found on:

http://ihcp.jrc.ec.europa.eu/our_labs/eurl-ecvam/laboratories-research/predictive_toxicology/qsar_tools/toxtree

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